# **Depth, hard coral, and turf cover as predictors of micro‑scale spatial distribution of fshes in a subtropical rocky reef**

**Thiago E. Rodrigue[s](http://orcid.org/0009-0001-3591-1096) · Vinícius J. Giglio · Fernando Z[. Gi](http://orcid.org/0000-0001-6078-2241)bran · Fernand[a A.](http://orcid.org/0000-0003-1285-6829) Rolim · Leonardo M. Neve[s](http://orcid.org/0000-0002-0269-6588) · [Gui](http://orcid.org/0000-0001-9134-595X)lherme Henriq[ue P](http://orcid.org/0000-0002-8009-8890)ereira‑Filho · Gustavo Shintate · Gabriel R. S. Souza · Domingos Garrone‑Neto · Fabio S. Motta**

Received: 21 June 2023 / Accepted: 15 August 2024 © The Author(s), under exclusive licence to Springer Nature B.V. 2024

**Abstract** The combination of the physical structural heterogeneity of the environment, oceanographic characteristics, and the benthic assemblage composition structures the habitat, consequently shaping the associated diversity of fsh. Understanding the spatial variability of fish assemblages and how it relates to environmental factors is essential to identify potential variables that determine spatial patterns and predict impacts on fsh assemblage metrics, thus providing valuable information for management. Here, we investigated reef fsh microscale spatial distribution around Alcatrazes Island in the subtropical region of the Southwest Atlantic, Brazil. Multivariate Regression Trees were ftted to explore the

contains supplementary material available at [https://doi.](https://doi.org/10.1007/s10641-024-01590-7) [org/10.1007/s10641-024-01590-7.](https://doi.org/10.1007/s10641-024-01590-7)

T. E. Rodrigues · V. J. Giglio · F. A. Rolim · G. H. Pereira-Filho  $\cdot$  G. Shintate  $\cdot$  F. S. Motta ( $\boxtimes$ ) Laboratório de Ecologia e Conservação Marinha (LABECMar), Instituto do Mar, Universidade Federal de São Paulo, R. Carvalho de Mendonça, 144, Santos, SP CEP 11070-102, Brazil e-mail: fmotta@unifesp.br

T. E. Rodrigues Serra Mar Ambiental. Rua Tupiniquins, 6, Coroa Vermelha, Santa Cruz Cabrália, BA, Brazil

V. J. Giglio Universidade Federal do Oeste do Pará, Campus Oriximiná, Rodovia PA 439, 257, Oriximiná CEP 68270-000, Brazil

efects of structural heterogeneity, wave power, depth, water temperature, and benthic cover on the structure of reef fsh assemblages, addressing composition, richness, density, biomass, trophic groups, mobility, and conservation status. Our results suggest that depth, turf, and coral cover were the main predictors of fish assemblages in rocky reefs at Alcatrazes Island, revealing five distinct fsh assemblages with diferent habitat preferences. These results provide additional insights into the relationship between reef fshes and their environment, providing empirical evidence for decision-makers to implement spatially based management policies, especially to prioritize zones for conservation along the island.

**Keywords** Reef fish  $\cdot$  Environmental variables  $\cdot$ **Supplementary Information** The online version Rocky reef · Benthic cover · Southwestern Atlantic

### F. Z. Gibran

Centro de Ciências Naturais e Humanas, Universidade Federal do ABC, Rua Arcturus, 03, Jardim Antares, São Bernardo do Campo, SP CEP 09606-070, Brazil

#### L. M. Neves

Departamento de Ciências do Meio Ambiente, Universidade Federal Rural do Rio de Janeiro, Av. Pref. Alberto da Silva Lavinas, Três Rios, RJ 184725802-100, Brazil

G. R. S. Souza · D. Garrone-Neto Laboratório de Ictiologia e Conservação de Peixes Neotropicais, Universidade Estadual Paulista, Campus de Registro, Avenida Saburo Kameyama 375, Agrochá, Registro, SP 11900-000, Brazil

# **Introduction**

Reef fish assemblages respond to changes in environmental conditions with fuctuations in richness and abundance at diferent spatial and temporal scales (Parravicini et al. [2013\)](#page-15-0). The dynamics of community structure in marine environments at the local level are infuenced by population relationships to biotic and habitat characteristics. In reef systems, biochemical and physical factors with the potential to shape biological assemblages include water temperature (Floeter et al. [2001;](#page-14-0) Bellwood et al. [2005](#page-13-0)), wave exposure (Friedlander et al. [2003;](#page-14-1) Maia et al. [2018](#page-15-1)), depth gradient (Teixeira-Neves et al. [2015](#page-16-0); Pereira et al. [2018](#page-15-2); Matheus et al. [2019](#page-15-3)), and nutrient production (Williams et al. [2015](#page-16-1)). Such environmental features shape the geomorphology, which infuences sessile benthic communities by afecting reef organism recruitment and settlement (Carleton and Sammarco [1987;](#page-13-1) Mallela [2018\)](#page-15-4), thus infuencing key structural features of fsh habitats (Sale [2013;](#page-16-2) Teixeira-Neves et al. [2015;](#page-16-0) Russ et al. [2021](#page-15-5)). For instance, increasing reef structural complexity has been described as beneficial to species richness by providing more space in quantity and types of shelters (e.g., Friedlander et al. [2003;](#page-14-1) Darling et al. [2017\)](#page-14-2). Therefore, the spatial context can infuence the structure of reef fsh assemblages, with relevant mechanisms involved directly related to the scale of analysis.

Identifying the variables infuencing the spatial distribution of reef fshes is important to understand the forces that shape assemblage structure, providing data to guide management strategies such as spatial planning (Haupt et al. [2017](#page-14-3)). Variations in local environmental conditions can infuence feeding, metabolism, growth rates, pelagic larvae duration, and even species survival (O'Connor et al. [2007](#page-15-6)). However, in many sites, disentangling the importance of environmental variables in the structure of fsh communities has been challenging when they occur synergistically with sources of anthropogenic impacts such as fshing and tourism (Richards et al. [2012\)](#page-15-7). To reduce the potential bias, studying sites with minimal human interference at the local level, such as efective and well-enforced no-take marine protected areas (MPAs), can be ideal.

Studies investigating spatial patterns of the ichthyofauna based on environmental predictors have been conducted at large (i.e., > 50 km; Bender et al. [2013](#page-13-2); Endo et al. [2019](#page-14-4); Quimbayo et al. [2019\)](#page-15-8) and small spatial scales (Friedlander and Parrish [1998](#page-14-5); Arias-González et al. [2006](#page-13-3); Brokovich et al. [2006;](#page-13-4) Cvitanovic and Hoey [2010;](#page-14-6) Agudo-Adriani et al. [2019\)](#page-13-5), with most studies focusing on tropical reefs. However, patterns of variation in reef fsh assemblages over micro-spatial scales (i.e., from meters to  $< 2$  km) in subtropical reefs remain less explored (García-Charton et al. [2004](#page-14-7); Teixeira-Neves et al. [2015](#page-16-0); Neves et al. [2016](#page-15-9); Ferrari et al. [2018](#page-14-8); Nanami [2022;](#page-15-10) Barreto et al. [2024](#page-13-6)).

In this study, we investigated the spatial distribution of reef fsh assemblages at Alcatrazes island, a subtropical coastal island off Brazil, Southwestern Atlantic. More specifcally, we explored the potential efects of structural heterogeneity, wave power, depth, water temperature, and benthic cover on reef fish assemblages' structure, composition, richness, density, biomass, trophic groups, mobility, and conservation status.

# **Material and methods**

#### Study area

Alcatrazes is the largest island of the Alcatrazes Archipelago, formed by six islands located 33 km away from the coast of the São Paulo state, Southeastern Brazil (Southwestern Atlantic; Fig. [1](#page-2-0)). Alcatrazes Island encompasses two no-take MPAs: the Tupinambás Ecological Station (ESEC–IUCN Category Ia), created in 1987, and the Alcatrazes Archipelago Wildlife Refuge (REVIS–IUCN Category III), established in 2016. The Archipelago features rich terrestrial and marine biodiversity with several terrestrial endemic and threatened species (ICMBio [2017](#page-14-9)). Regarding oceanographic conditions, the region is infuenced by three water masses transported by the Brazil Current, one of which, the South Atlantic Coastal Water, is associated with coastal upwelling (Castro et al. [2008](#page-13-7)). The upwelling occurs mainly in the austral summer, changing the physicochemical parameters with consequences for biological processes in the region (Pires-Vanin et al. [1993](#page-15-11)).

The sublittoral rocky bottom is characterized by its range of structural features, from fat plateaus of large and uniform rock walls (exhibiting little to no structural complexity), to large boulders up to 15 m in diameter, spaced a few meters apart, and featuring numerous



<span id="page-2-0"></span>**Fig. 1** The study area is located within the Wildlife Refuge of the Alcatrazes Archipelago (blue polygon), on the north coast of the State of São Paulo. The green circles represent the Tupinambás Ecological Station, a no-take marine protected area created in 1986, covering areas of the Alcatrazes Archipelago

crevices and holes. The leeward side of the island, with less hydrodynamic energy, harbors the richest benthic sessile cover, while encrusting calcareous algae dominate the windward sites, susceptible to waves and currents, with lower habitat complexity and lower benthic cover richness (Gibran and Moura [2012](#page-14-10)).

# Data collection

The present study was authorized by the Brazilian Ministry of Environment through the Brazilian System of Information and Authorization in Biodiversity (SISBIO No. 46206–8) and approved by the Federal University of São Paulo Ethics Committee (CEUA No. 1015170420).

(south sector 1) and Palmas and Cabras Islands (north sector 2). The detailed map shows Alcatrazes Island with the eight sample sites. 1 Matacões, 2 Raia, 3 Baba de Boi, 4 Tartaruga, 5 Paredão, 6 Oratório, 7 Geladeira, and 8 Jardim dos Corais

#### *Reef fsh assemblages*

Data on reef fsh abundance was collected around Alcatrazes Island in October 2018 (during the South Hemisphere Spring season), through 218 stationary visual censuses using scuba diving (adapted from Minte-Vera et al. [2008](#page-15-12)) at eight reef sites (Baba de Boi *n*=27; Geladeira *n*=35, Jardim dos Corais *n*=34; Matacões *n*=30; Paredão *n*=26; Raia *n*=25; Saco do Oratório *n*=30; Tartaruga, *n*=11). All sites, separated by at least 300 m, were diferentiated and selected during the monitoring of the archipelago based on the range of habitats found, primarily infuenced by topography and benthic composition. Therefore, the main aim of this study was to investigate the

variations in fsh assemblages across these distinct sites to inform management decisions.

The abundance of every fsh species and individual body size were estimated within an observer-centered cylinder with 4-m radius for 5 min. Individuals with total length (TL) smaller than 10 cm were counted within a nested cylinder with a 2-m radius in two size classes ( $\leq$ 2 cm and 2–10 cm), while individu $als > 10$  cm TL were counted within a 4-m radius in four size classes (10–20; 20–30; 30–40; and>40 cm). Stationary visual census has been historically used as a method for monitoring this no-take MPA (Motta et al. [2021](#page-15-13); Rolim et al. [2024](#page-15-14)), as well as other marine reserves in Brazil (e.g., Moura and Francini-Filho [2005;](#page-15-15) Freitas et al. [2019\)](#page-14-11). They are particularly useful for large reefs with high structural heterogeneity with diverse and abundant fish populations (Bohnsack and Bannerot [1986](#page-13-8); Minte-Vera et al. [2008\)](#page-15-12). Sampling was conducted between 9 a.m. and 5 p.m. BRT (GMT–3) to avoid bias due to changes in fsh activity between day and night periods (Hobson [1965](#page-14-12); Amaral et al. [2023\)](#page-13-9). The minimum horizontal transparency of the seawater during data collection was 5 (i.e., larger than the sample unit radius).

# *Environmental variables*

Structural features of the reef were examined to assign a broad structural complexity score for each fish sampling, following Wilson et al.  $(2007)$  $(2007)$ . The scores were 1 for the lowest, 2 for intermediate, and 3 for the highest structural complexity. Flat substrates with low potential to provide shelter for the reef biota receive scores of 1, whereas those with rocks, holes, and crevices with homogenous sizes receive scores of 2. Reefs with high complexity with many rocks, holes, and crevices of diferent sizes were assigned score 3. The standard deviation of structural complexity was calculated based on the scores averaged across all sampling units at each site (eight levels). This average served as a measure of heterogeneity, with higher heterogeneity values indicating sites with greater variation in structural complexity.

Data on wave power was gathered from Takase et al. ([2021\)](#page-16-4). These authors investigated the level of exposure to wave power distribution around Alcatrazes Island by applying a numerical model from a 14-year time series (2005–2018) extracted from the Global Wave Watch III model. Wave power was estimated through the relation that considers the synergistic efect between wave height and period (Takase et al. [2021\)](#page-16-4). Data on wave power were extracted during springs (average for years 2005–2018) for each site (eight levels), the same season we sampled reef fsh assemblages in 2018.

Depth and water temperature were collected in each sampling unit, i.e., census (218 levels) through a dive computer (Puck Pro Model–Mares®) with precision levels of  $\pm$  0.2 m and  $\pm$  2 °C, respectively. During data collection, depths ranged from 2 to 24 m and the water temperature ranged from 20 to 25 °C.

The characterization of the benthic cover was conducted through 10 photoquadrats of  $70\times70$  cm distributed randomly in each sampling site. Each photoquadrat was composed of a mosaic of 15 digital images. The percentage cover was estimated using the software photoQuad (Trygonis and Sini [2012\)](#page-16-5). In this software, twenty randomly located points were generated per digital image, thus totaling 300 points per photoquadrat. The organisms or colonies immediately below each point were classifed into 12 major benthic categories: sand, bare rock, crustose coralline algae (CCA), frondose algae, coral, tunicate, soft coral (Zoantharia), cenocytic algae, Octocorallia, Echinodermata, Porifera, bivalve, Hydrozoa, and turf algae. Subsequently, the average of the percentage cover of each benthic category was calculated by site  $(n=8)$ . The dominant broad benthic categories  $(5\%)$  were selected for the analysis, resulting in turf, frondose algae, soft corals (i.e., *Carijoa riisei* and *Palythoa caribaeorum*), and corals (i.e., *Madracis decactis* and *Mussismilia hispida*).

# Data analysis

Fish species were categorized according to Pinheiro et al. ([2018\)](#page-15-16) in trophic groups as herbivores (HERB), macrocarnivores (MCAR), mobile invertebrate feeders (MINV), sessile invertebrate feeders (SINV), omnivores (OMNI), and planktivores (PLANK), as well as in mobility as highly mobile or migratory (HMO), roving (ROV), and sedentary or territorial (SED). The species were also assigned a conservation status based on the International Union for the Conservation of Nature Red List and the Brazilian Red List, as either Least Concern (LC), Near Threatened (NT), or Vulnerable (VU) (IUCN [2023](#page-14-13)). Fish biomass was estimated using length–weight relationships available in FishBase (Froese and Pauly [2020](#page-14-14)). The calculations were done using the midpoint values for each length category (i.e., individuals in the 10–20 cm class were calculated as measuring 15 cm). When coefficient values were not available for a species, we used coefficients of the closest related species or genus. For each species, the relative abundance and relative biomass were calculated by dividing its number of individuals and total biomass by the corresponding totals for all species in the community. Mean density and biomass were then determined by dividing the species-specifc abundance and biomass by the census area, yielding data in individuals and grams per square meter.

Multivariate regression trees (MRT) were ftted for the fish census data  $(n=218$  levels) using the R package *mvpart* to explore the infuence of environmental variables (standard deviation of structural heterogeneity, wave power, depth, water temperature, and benthic cover) on the spatial distribution of reef fshes considering species as taxonomic level. The MRT allows recursive partitioning of quantitative or categorical environmental variables, determining the formation of clusters. The calculation occurs from repeated divisions of diferent sample groups, producing "nodes" as homogenous as possible concerning the response variable. Homogeneity is measured by fnding the best division that minimizes the sums of squares over the multivariate average within each node (De'Ath [2002](#page-14-15)). The best tree was selected by choosing the highest cross-validation value and lowest value of the relative cross-validated error (De'Ath [2002\)](#page-14-15). For analysis purposes, each category of benthic organism was added as an independent variable in the MRT.

Hellinger's pre-transformation was used on the species matrix, where the abundance values of each fish species are divided by the total abundance of the site. To explore patterns in an assemblage's segregation according to environmental variables, a non-metric multidimensional scaling (nMDS) over Euclidian dissimilarity matrices was calculated. This procedure allowed us to corroborate the explanatory potential of the outputs from the tree generated by the MRT. Diferences in richness and biomass between assemblages revealed by the MRT were tested using the Kruskal–Wallis test and the post hoc Dunn test.

Boosted regression trees (BRT) were ftted for the census data (218 levels) using the R package *gbm* (Greenwell et al. [2019\)](#page-14-16) to understand the drivers of the abundance of the most important fish species that

contributed to splitting assemblages revealed by the MRT. Boosted regression tree (BRT) models are a modeling approach whereby a succession of regression trees is developed using machine learning models. To avoid model overftting and attain the highest accuracy, as indicated by the lowest values of cross-validation deviance and standard error, optimal BRT models were selected by examining all possible combinations of values for bag fraction (0.5 and 0.75), learning rate (0.001, 0.005, 0.01, and 0.05), and tree complexity (1 to 5). BRT procedures were conducted following the protocol described by Elith et al. [\(2008](#page-14-17)). All analyses were conducted in the R software version 3.0.3 (R Core Team [2011](#page-15-17)).

# **Results**

Fish assemblage structure and conservation aspects

We recorded a total of 14,524 individuals of 75 species, 49 genera, and 31 families at Alcatrazes Island. The ten most abundant fish species accounted for 88.9% of total fsh abundance and for 69.5% of biomass (Table [1\)](#page-5-0). The most common species, *Haemulon aurolineatum*, was found in 88.7% of censuses, representing 60.9% of the total abundance and 17.8% of the sampled biomass (Table [1\)](#page-5-0). Haemulidae was the family with the largest abundance, followed by Pomacentridae, Holocentridae, Kyphosidae, Labridae, and the others (Fig. [2A](#page-6-0)). The most speciose families were Labridae with 12 species, Pomacentridae with seven species, followed by Carangidae and Haemulidae with six species each. The average richness per sampling was  $8.3 \pm 3.6 \ (\pm s.d.)$  with a maximum of 19 species. The average abundance of specimens per sampling was  $65.7 \pm 57.7$ , ranging from 0 to 377 fshes. The average biomass per census was  $112.5 \pm 125.5$  g/m<sup>2</sup>, ranging from 0 to 804 g/m<sup>2</sup>.

Regarding trophic groups, most fsh sampled were mobile invertebrate feeders, accounting for 74.3% of all individuals counted and for 46.5% of total biomass (Fig. [2B](#page-6-0)), followed by omnivores (11.3% of biomass), herbivores (10.5%), planktivores (1.4%), sessile invertebrate feeders (1.4%), and macrocarnivores (1.1%). Mobile invertebrate feeders were the most speciose group, with 26 species, followed by macrocarnivores  $(n=18; Fig. 2B)$  $(n=18; Fig. 2B)$  $(n=18; Fig. 2B)$ .

Seven fish species recorded in Alcatrazes Island are listed in the Brazilian Red List of Threatened Species

<span id="page-5-0"></span>

ebrate feeders (SINV), omnivores (OMNI), and planktivores PLANK). The full table with all species sampled is provided Table S1



as Vulnerable. Three threatened species were macrocarnivores of the Epinephelidae family (i.e., *Epinephelus marginatus*, *Mycteroperca bonaci*, and *Epinephelus morio*), where *Epinephelus marginatus* was the most abundant of the three species. Three species were herbivores of the Labridae family (*Sparisoma axillare*, *Sparisoma frondosum*, and *Scarus zelindae*). One species was a mobile invertebrate feeder, *Elacatinus fgaro*. Threatened species represented 1.64% of total fsh abundance and 8.26% of fsh biomass.

Infuence of environmental variables on the spatial distribution of reef fsh

The most parsimonious MRT model, presenting greater explanatory power, lower error, and low number of nodes, was selected, resulting in fve distinct assemblages (Supplementary Fig. 1; Fig. [3](#page-7-0)A,B). The MRT model selected explained 18.9% of data variability, with a cross-validated relative error (CVRE) of 0.88.

Three explanatory variables (depth, coral cover, and turf) were identifed as the most important factors infuencing the spatial distribution of reef fsh assemblages. The nMDS ordination shows a clear overlap among three reef fish assemblages (Fig. [3C](#page-7-0)).

Assemblage A represented 79 censuses (Fig. [3B](#page-7-0); Table [2](#page-7-1)), being dominated by the mobile invertebrate feeder *Haemulon aurolineatum* (33.2±34.2 ind.  $census^{-1}$ ), but the discriminating species was the omnivore *Diplodus argenteus* (2.8±3.3 ind. census−1). In this assemblage, areas of intermediate depth (6.45 m and 10.75 m) were grouped into fve sites with coral coverage of less than 4.16% (Fig. [3A](#page-7-0), B).

Assemblage B (8 censuses) was represented by the mobile invertebrate feeder *Coryphopterus glaucofraenum* (7 $\pm$ 4 ind. census<sup>-1</sup>), but the discriminating species was the herbivore *Stegastes fuscus*  $(2.25 \pm 1.83)$ ind. census−1) which was related to greater coral cover ( $\geq$ 4.16%) at depths greater than 10.75 m and grouped into samples from two sites.

Assemblage C (27 censuses) was most represented by *Haemulon aurolineatum*  $(63.2 \pm 56.8 \text{ ind.})$ census<sup> $-1$ </sup>). This assemblage was driven by a higher coral cover ( $\geq$  4.16%) and intermediate depths ( $\geq$  6.45 to  $<$  10.75 m). Here, samples from three sites were grouped (Fig. [3A](#page-7-0)).

The two remaining assemblages, D and E, featured the omnivore *Abudefduf saxatilis* and the herbivore *Kyphosus* spp. as discriminator species, the frst species contributing with greater abundance and the second with greater biomass.

Assemblage D (six censuses) was most represented by the omnivore *Abudefduf saxatilis*  $(27.8 \pm 10.5$ ind. census<sup> $-1$ </sup>), at depths shallower than 6.45 m, and higher turf cover  $(\geq 82.84\%)$ . Here, samples were found only in one site, the most windward location, called "Paredão" (Fig. [3](#page-7-0)A, B).

Assemblage E (98 censuses) was dominated by *Haemulon aurolineatum*  $(50.3 \pm 58.8 \text{ ind.} \text{ census}^{-1})$ , followed by *Abudefduf saxatilis* (15.3±12.8 ind. census−1). It was characteristic for shallower reefs  $(<6.45$  m) and turf cover lower than 82.84%. Here, samples from shallower reefs of seven sites were grouped.

<span id="page-6-0"></span>**Fig. 2** Abundance by family (**A**) and trophic group (**B**). Numbers associated with each bar indicate total species richness within each fish family or trophic group. Trophic groups: mobile invertebrate feeders (MINV), omnivores (OMNI), herbivores (HERB), planktivores (PLANK), sessile invertebrate feeders (SINV), and macrocarnivores (MCAR)





<span id="page-7-0"></span>**Fig. 3 A** Sampled sites (numbers) around Alcatrazes Island and fish assemblages (colored circles) revealed by the multivariate regression tree (MRT). **B** MRT showing reef fsh assemblages on Alcatrazes Island according to environmental variables: depth, hard coral cover, and turf. The variables are divided by the nodes (black circles), and the discriminating fish species responsible for the node and their relative contribution from the division to the total explained variance of the

model are given at the bottom. Discriminant species: *ABUSAX*, *Abudefduf saxatilis*; *KYPSP*, *Kyphosus* spp.; *DIPARG*, *Diplodus argenteus*; *HAEAUR*, *Haemulon aurolineatum*; and *STEFUS*, *Stegastes fuscus*. "*N*" is the number of corresponding censuses for each assemblage distinguished by the MRT (colored circles labeled A, B, C, D, and E). **C** Principal components analysis (PCA) of fve assemblages revealed by the MRT, illustrated through diferent colors

<span id="page-7-1"></span>**Table 2** Description (richness and biomass density) of each assemblage revealed by the multiple regression tree model, in terms of number of censuses in each assemblage (*N*), as well as minimum (Min), mean, standard deviation (Std Dev), median, and maximum (Max) values



Concerning assemblages' richness, the assemblages C and E presented the higher values, but were not signifcantly diferent from assemblage D (Fig. [4A](#page-8-0), Table [2](#page-7-1)). The lowest richness values were registered in assemblages A and B (Fig. [4](#page-8-0)A, Table [2](#page-7-1)), but not signifcantly diferent from D. In terms of biomass  $(g/m^2)$ , assemblage B presented significantly lower values (Fig. [4B](#page-8-0), Table [2\)](#page-7-1). Assemblages A, C, and D were not signifcantly diferent and assemblage E presented the highest biomass values, although not difering from assemblage D (Fig. [4](#page-8-0)B, Table [2](#page-7-1)).

The composition of trophic groups varied among assemblages, indicating a higher density of mobile invertebrate feeders in the deeper assemblages A, B, and C. The assemblages D and E, located in shallower areas, had a higher omnivore density, particularly D, which is exclusive to the reef area more exposed to wave action (Fig. [5](#page-9-0)). Despite showing similar density,



<span id="page-8-0"></span>**Fig. 4 A** Reef fsh richness (number of species) by census. **B** Total biomass  $(g/m^2)$  of the sampled censuses in each of the five assemblages revealed by the MRT model. The violin shape represents the distribution of data, the boxplot within the violins represents frst and third quartiles of the data, vertical lines are the 95% inferior and superior limits, and the dots are the outliers. Letters above assemblages represent diferences between groups detected by a Dunn test  $(P < 0.05)$ 

the biomass of sessile invertebrate feeders and macrocarnivores was higher in deeper assemblages, while shallower assemblages displayed higher biomass of herbivores and omnivores (Fig. [5\)](#page-9-0).

Regarding mobility, assemblages were dominated by roving species. Only assemblage B featured a higher density of sedentary or territorial species (78%), composed mainly by *Coryphopterus glaucofraenum* and *Stegastes fuscus*, but most of the biomass was of roving species (85%; Fig. [5;](#page-9-0) Supplementary Table 1).

Within the assemblages, according to the Brazilian Red List of Threatened Species, assemblage B comprised a higher density of Vulnerable species, represented mainly by *Elacatinus fgaro* and *Epinephelus marginatus*. Assemblage E featured greater biomass of the Vulnerable species *Mycteroperca bonaci* and *Sparisoma frondosum*. Regarding biomass, the assemblages B and E comprised a higher amount of Vulnerable and Near Threatened species, representing 11% and 15% of biomass, respectively.

The BRT models corroborated patterns observed in the MRT and add more insights on the main environmental drivers shaping the spatial distribution of the most abundant species of reef fshes (Fig. [6](#page-10-0)). Depth was the most important predictor for *Abudefduf saxatilis* and *Kyphosus* spp., which were most abundant at depths<6.45 m. *Stegastes fuscus* was more abundant at low turf cover and at depths between  $\geq$  6.45 and <10.75 m. Higher abundances of *Haemulon aurolineatum* were mostly correlated with high coral cover and low turf, with the opposite being observed for *Diplodus argenteus* (Fig. [6](#page-10-0)).

# **Discussion**

Our results revealed that the combination of depth, turf, and coral cover is an important predictor of fish assemblages at a micro-scale in the subtropical reefs of Alcatrazes Island, in the Southwestern Atlantic Ocean. The analysis revealed fve distinctive fish assemblages with different habitat preferences around the island.

Fish assemblages' structure in a regional context

The reef fsh fauna recorded in Alcatrazes Island, comprising 75 species, represents approximately 10% of the total reef fsh fauna found in the



<span id="page-9-0"></span>**Fig. 5** Relative percentage (%) of density (left column) and biomass (right column) of reef fshes in the fve assemblages revealed by the MRT—letters A, B, C, D, and E at the bottom of the fgure. The *y*-axis of each subplot ranges from 0 to 100% in terms of relative percentage. Trophic groups: MCAR, macrocarnivores; MINV, mobile invertebrate feeder; HERB,

Brazilian province (733 species) and 18% of resident reef fshes (73 out of 405 species) according to Pinheiro et al.  $(2018)$ . This region encompasses the east-southeast subprovince, characterized by a transition from biogenic reefs to rocky reefs (Floeter et al. [2001;](#page-14-0) Pinheiro et al. [2018\)](#page-15-16). It serves as an area of convergence for species from both environments (Floeter et al. [2001;](#page-14-0) Pinheiro et al. [2018\)](#page-15-16). Indeed, the species list compiled in this study includes representatives from tropical waters, such as haemulids and labrids, and temperate waters, i.e., serranids and scarids (Supplementary Table 1, Floeter et al. [2001](#page-14-0); Pinheiro et al. [2018\)](#page-15-16).

The overall reef fsh richness and trophic group composition in Alcatrazes Island resembled those of reefs in nearby coastal islands, as indicated by

herbivore; OMNI, omnivores; PLANK, planktivore; and SINV, sessile invertebrate feeder. Mobility HMO, highly mobile or migratory; ROV, roving; and SED, sedentary or territorial. Conservation status: LC, Least Concern; NT, Near Threatened; and VU, Vulnerable

studies conducted by Souza et al. [\(2018\)](#page-16-6) and Rolim et al. ([2019](#page-15-18)), but were higher than those observed in sites closer to shore (Gibran and Moura [2012](#page-14-10)). This discrepancy underscores the importance of geographical factors, such as distance from the coast, in shaping reef fsh assemblages in the region (Teixeira-Neves et al. [2015](#page-16-0); Silva et al. [2021\)](#page-16-7). Furthermore, coastal sites experience greater pressure from both legal and illegal human activities, including fshing (Imoto et al. [2016](#page-14-18)) and pollution (von Glasow et al. [2013\)](#page-16-8) especially near megacities, due to their accessibility, highlighting the infuence of sociocultural aspects on reef fsh dynamics (Mora et al. [2011\)](#page-15-19). The interplay between geographical and social characteristics likely contributes to this crossshelf gradient in reef fsh richness.



<span id="page-10-0"></span>Fig. 6 Partial dependence plots revealed by the boosted regression trees using the three main predictors (depth, hard coral cover, and turf cover) and abundance of reef fsh species characteristic for the assemblage revealed by the MRT

Predictors of spatial patterns in rocky reef fish assemblages

In the present study, depth was an important predictor for total biomass, in which the lowest values were registered in the deepest assemblage  $(>10.75 \text{ m})$ , as well as for some species, such as *Abudefduf saxatilis* and *Kyphosus* spp., whose abundances decreased with depth. Depth has been demonstrated as an important environmental variable afecting the structure of reef fsh assemblages at a local level (Francini-Filho and Moura [2008](#page-14-19); Pereira-Filho et al. [2011](#page-15-20); Gibran and

Moura [2012](#page-14-10); Pereira et al. [2018;](#page-15-2) Silva et al. [2021](#page-16-7)). Depth may infuence reef fsh assemblages by altering physical variables (e.g., wave exposure, temperature, luminosity, and salinity) and by modifying benthic assemblages' composition and structural complexity (García-Charton et al. [2004;](#page-14-7) Luiz et al. [2015;](#page-15-21) Ferrari et al. [2018](#page-14-8)). At Alcatrazes Island, deeper reefs usually present lower complexity and lower temperatures (Gibran and Moura [2012](#page-14-10)), which may limit the abundance of species with a preference for more complex shallow warm waters.

Intermediate depths seem to favor the territorialist species *Stegastes fuscus*, probably by providing light and exposure in ideal conditions for their food to grow, which is mainly composed of red flamentous algae (Ferreira et al. [1998\)](#page-14-20). Turf cover was also an important predictor for the species, with higher abundances associated with less turf. This is not expected, since turfs are composed of multiple species of algae, usually including flamentous algae (Ferreira et al. [1998\)](#page-14-20). As turf in the present study was not analyzed carefully, the composition of the algae species may change across habitats and depths along the reef. This topic needs further investigation concerning the diferent compositions of the turf in the Alcatrazes Island to properly investigate the relationship with *Stegastes fuscus* abundance.

Wave exposure is known to infuence reef fsh vertical distribution. More exposed reefs usually harbor lower richness and abundance of species (Luiz et al. [2015\)](#page-15-21). In Alcatrazes Island, only one reef sampled was on the windward side exposed to a higher wave power, and the assemblage D occurred exclusively at this site, being predicted by shallow depths and higher turf cover. Despite the more extreme wave conditions, this assemblage did not show signifcant diferences in species richness and biomass compared to leeward sites, yet it constituted a distinct assemblage specialized for such an environment. Assemblage D was characterized by a higher density of the omnivore species *Abudefduf saxatilis*, followed by *Diplodus argenteus* and *Anisotremus virginicus*. Such species have a diversifed diet and do not present a well-established distribution pattern in the rocky shore habitats (Luiz et al. [2008](#page-15-22); Anderson et al. [2015](#page-13-10)). With increased levels of exposure, the composition of benthic assemblages is also infuenced, leading to the selection of species more resistant to challenging conditions

(Matheus et al. [2019\)](#page-15-3). Consequently, it is expected that higher abundances of fsh species with less specialized resource demands will be recorded in such environments.

Regardless of the combination of predicted environmental variables, *Haemulon aurolineatum* was the most abundant species associated with higher coral cover  $(\geq 4.16\%)$ . In tropical coral reefs, the species is associated with middle-depth sites and intermediate to higher levels of coral cover (Francini-Filho and Moura [2008\)](#page-14-19). However, at Alcatrazes Island, depth appears to not be a predictor of this species' occurrence, which was more related to benthic cover. *Haemulon aurolineatum* is dominant in subtropical rocky shores of the Western Atlantic, inhabiting reef slopes, reef-sand interfaces, and sandy bottoms, across a range of depths, and usually corresponding to more than 50% of fsh species abundance (Luiz et al. [2008;](#page-15-22) Daros et al. [2012](#page-14-21); Anderson et al. [2020\)](#page-13-11). Juveniles of this species form large schools with more than 100 individuals and are associated with the reef at all life stages (Quimbayo et al. [2021](#page-15-23)). The preference for areas with high coral cover is likely related to foraging activities, with juveniles (<10 cm) exploring food resources suspended in the water column above the reef, mostly copepods and amphipods, while adults  $(>15 \text{ cm})$  primarily forage in the sand around the reef, usually feeding on polychaetes and crabs (Pereira and Ferreira [2013](#page-15-24); Pereira et al. [2015\)](#page-15-25). This close association with reef environments elucidates the positive efect of coral cover on the abundance of *Haemulon aurolineatum* at Alcatrazes Island, without a corresponding depth efect. The absence of a depth effect suggests that the species' foraging behavior is predominantly infuenced by the availability of resources within reef habitats.

Turf algae, representing a multitude of short turfforming species of macroalgae belonging to the orders Corallinales, Ceramiales, and other green and reef flamentous algae, is the dominant group of benthic organisms in most subtropical reefs (Aued et al. [2018\)](#page-13-12). These communities play key roles in the ecology and trophodynamics of reef systems, occupying substantial areas of the reef from shallow habitats to the interface with soft bottoms in deeper reefs. At Alcatrazes Island, the distribution of *D. argenteus* was related to turf cover, a pattern observed in other subtropical rocky shores of the Brazilian province (Teixeira-Neves et al. [2015](#page-16-0)). Turf is usually not related to increased structural reef complexity, but it does support a diverse and abundant associated macrofauna that lives in the turf matrix and detritus, being an important resource for invertivore and herbivore species (Ferreira et al. [2004;](#page-14-22) Liuzzi and Gappa [2008](#page-14-23)).

The MRT model accounted for 18.9% of the variability in the data, and supplementary analytical approaches (PCA and BRT) supported the diferentiation of reef fsh assemblages and the signifcance of environmental variables in predicting reef fsh fauna composition. We propose that variables beyond those we investigated may serve as predictors of reef fish assemblages on rocky shores, such as biotic factors like primary productivity in the water column. The abundance of plankton can significantly impact the distribution and abundance of reef fsh species. According to the Resource Availability Theory, fuctuations in resource availability can infuence the intensity of resource competition, potentially leading to evolutionary changes in life-history traits (Endara and Coley, 2011). Plankton constitute a crucial food source during the juvenile stages of certain reef fish species, such as *Haemulon* spp. and *Kyphosus* spp., which undergo ontogenetic shifts in their diet. While younger individuals predominantly consume planktonic items in the water column, adult *Haemulon* spp. favor invertebrates associated with reefs, and *Kyphosus* spp. specialize as herbivores (Silvano and Güth [2006;](#page-16-9) Pereira et al. [2015\)](#page-15-25).

In addition, assemblages may vary temporally, infuenced by seasonal wave exposure and water masses waters that cause a seasonal thermocline in the region. Further investigation is encouraged to verify if the spatial variation changes temporally (i.e., spatiotemporal patterns). Temporal changes in fsh abundance can be relevant, for instance, after recruitment pulses (Lewis [1997](#page-14-24)), or under the infuence of seasonal upwelling (Cordeiro et al.  $2016$ ), or communities may even remain spatially consistent over time (Anderson and Millar [2004](#page-13-13); Sánchez-Caballero et al. [2019](#page-16-10)).

Anthropogenic drivers have been increasingly described as predictors of patterns in biodiversity. For instance, fishing has changed fish assemblages in subtropical Brazilian reefs by reducing the density of macrocarnivores and herbivores (Anderson et al. [2014;](#page-13-14) Bender et al. [2014](#page-13-15)). Sites with increased protection levels and distant from the coast or in remote areas contain higher species richness and biomass (Quimbayo et al. [2019;](#page-15-8) Fonseca et al. [2021\)](#page-14-26). Alcatrazes Island was used by the Brazilian Navy for over two decades for military training, which has possibly decreased the occurrence of fshing activities in the region, but with insufficient enforcement. Since 2015, the MPAs of Alcatrazes Archipelago have increased enforcement levels compared to most Brazilian MPAs. Outcomes of the protection were a high reef fsh abundance compared to other coastal sites (Morais et al. [2017;](#page-15-26) Motta et al. [2021](#page-15-13)) and an increase in body size and abundance of fsheries target species of groupers and jacks compared to nearby fshed areas (Rolim et al. [2019;](#page-15-18) Motta et al. [2021\)](#page-15-13). The enhancement in management effectiveness (Giglio et al. [2019](#page-14-27)) and enforcement in marine reserves of the Alcatrazes Archipelago are expected to increase the conservation efects for reef communities. Since 2018, part of the Alcatrazes Archipelago is open for a guided visitation through diving and, more recently, boating through a strict code of conduct (Marconi et al. [2020;](#page-15-27) Giglio et al. [2022](#page-14-28)). Systematic monitoring was implemented to evaluate the potential efects of human presence on rocky reef biota. Such data has been used in the adaptive management of public use in Alcatrazes Wildlife Refuge (Giglio et al. [2022\)](#page-14-28). The continuation of the monitoring is essential to inform adaptive management on possible shifts in reef fsh assemblages from local and regional anthropogenic activities. In addition, efective management of partially protected MPAs surrounding the Alcatrazes Archipelago is important to ensure its efectiveness since fsh biomass inside marine reserves declines with increasing human impacts outside of reserves (Cinner et al. [2018\)](#page-14-29).

At the same site, two or more assemblages were structured by depth and environmental characteristics. Assemblages in shallow reefs were more homogenous, composed of assemblage D and E, while at depths greater than 6.45 m, assemblages A and C were the most common. Interestingly, on reefs located in the no-entry area inside the Tupinambás Ecological Station ("Saco do Oratório" and "Paredão") and border sites ("Geladeira" and "Tartaruga"), reef fsh assemblages were structured diferently. These sites harbor the highest coral cover at Alcatrazes Island and a higher density and biomass of threatened and sedentary species. Such fndings highlight the importance and suitability of the spatial zonation of use in Alcatrazes Island MPAs because the assemblages with distinct and important attributes are in the most protected area and at its border.

# **Conclusion**

Depth, coral cover, and turf cover were important predictors of the spatial distribution of resident fsh assemblages in subtropical rocky reefs of the Alcatrazes Island. Our results showed that two or more distinct fsh assemblages can be identifed in the same site in rocky reefs, varying according to local attributes. The outcomes of our study provide baseline data that can be useful for conservation use through spatial management approaches, considering local reef fish assemblage variation at local scales shaped by an assembly of processes, which is particularly important to improve the efectiveness of management strategies.

**Acknowledgements** We thank the Instituto Chico Mendes de Conservação da Biodiversidade (ICMBio) and the staff of Alcatrazes Wildlife Refuge and Tamoios Ecological Station for survey permits (SISBIO #62932–1). We thank the crews of "Netuno" and "Guardião" for their support in the feld. FS Motta and GH Pereira-Filho acknowledge individual grants from the Brazilian Research Council (CNPq).

**Funding** This study was funded by Instituto Linha D'água. The author Vinicius J Giglio received postdoctoral grant #2017/22273–0 from the São Paulo Research Foundation (FAPESP). The corresponding author Fabio S. Motta received individual grant # 2019/19423–5 from FAPESP. The authors are grateful to Instituto Linha D'água for essential fnancial support for monitoring expeditions.

**Data availability** The datasets generated during and/or analyzed during the current study are available from the corresponding author on reasonable request.

# **Declarations**

**Competing interests** The authors declare no competing interests.

# **References**

- <span id="page-13-5"></span>Agudo-Adriani EA, Cappelletto J, Cavada-Blanco F, Cróquer A (2019) Structural complexity and benthic cover explain reef-scale variability of fsh assemblages in Los Roques National Park. Venezuela Front Mar Sci 6:690. [https://doi.](https://doi.org/10.3389/fmars.2019.00690) [org/10.3389/fmars.2019.00690](https://doi.org/10.3389/fmars.2019.00690)
- <span id="page-13-9"></span>Amaral LSS, Bastos ASA, Carvalho-Júnior L, Maciel MDR, Teixeira-Neves TP, Araújo FG, Neves LM (2023) Diel changes in fish assemblages of Southwest Atlantic rocky reefs. Environ Biol Fish 106:627–639. [https://doi.org/10.](https://doi.org/10.1007/s10641-023-01400-6) [1007/s10641-023-01400-6](https://doi.org/10.1007/s10641-023-01400-6)
- <span id="page-13-13"></span>Anderson MJ, Millar RB (2004) Spatial variation and efects of habitat on temperate reef fsh assemblages in northeastern

New Zealand. J Exp Mar Bio Ecol 305:191–221. [https://](https://doi.org/10.1016/j.jembe.2003.12.011) [doi.org/10.1016/j.jembe.2003.12.011](https://doi.org/10.1016/j.jembe.2003.12.011)

- <span id="page-13-14"></span>Anderson AB, Bonaldo R, Barneche D et al (2014) Recovery of grouper assemblages indicates efectiveness of a marine protected area in Southern Brazil. Mar Ecol Prog Ser 514:207–215.<https://doi.org/10.3354/meps11032>
- <span id="page-13-10"></span>Anderson AB, Carvalho-Filho A, Morais RA et al (2015) Brazilian tropical fshes in their southern limit of distribution: checklist of Santa Catarina's rocky reef ichthyofauna, remarks and new records. Check List 11:1688. [https://doi.](https://doi.org/10.15560/11.4.1688) [org/10.15560/11.4.1688](https://doi.org/10.15560/11.4.1688)
- <span id="page-13-11"></span>Anderson AB, Joyeux JC, Floeter SR (2020) Spatiotemporal variations in density and biomass of rocky reef fsh in a biogeographic climatic transition zone: trends over 9 years, inside and outside the only nearshore no-take marine-protected area on the southern Brazilian coast. J Fish Biol 97:845–859.<https://doi.org/10.1111/jfb.14441>
- <span id="page-13-3"></span>Arias-González JE, Done TJ, Page CA et al (2006) Towards a reefscape ecology: relating biomass and trophic structure of fsh assemblages to habitat at Davies Reef, Australia. Mar Ecol Prog Ser 320:29–41. [https://doi.org/10.3354/](https://doi.org/10.3354/meps320029) [meps320029](https://doi.org/10.3354/meps320029)
- <span id="page-13-12"></span>Aued AW, Smith F, Quimbayo JP et al (2018) Large-scale patterns of benthic marine communities in the Brazilian Province. PLoS ONE 13:e0198452. [https://doi.org/10.](https://doi.org/10.1371/journal.pone.0198452) [1371/journal.pone.0198452](https://doi.org/10.1371/journal.pone.0198452)
- <span id="page-13-6"></span>Barreto CR, Quimbayo JP, Mendes TC, Cordeiro CAMM, Flores AAV (2024) Spatial structure and potential processes linking fish and benthic communities in a protected reef ecosystem in SE Brazil. Mar Ecol Prog Ser 738:41–55. <https://doi.org/10.3354/meps14589>
- <span id="page-13-0"></span>Bellwood DR, Hughes T, Connolly S, Tanner J (2005) Environmental and geometric constraints on Indo-Pacifc coral reef biodiversity. Ecol Lett 8:643–651. [https://doi.org/10.](https://doi.org/10.1111/j.1461-0248.2005.00763.x) [1111/j.1461-0248.2005.00763.x](https://doi.org/10.1111/j.1461-0248.2005.00763.x)
- <span id="page-13-2"></span>Bender MG, Pie MR, Rezende EL et al (2013) Biogeographic, historical and environmental infuences on the taxonomic and functional structure of Atlantic reef fsh assemblages. Glob Ecol Biogeogr 22:1173–1182. [https://doi.org/10.](https://doi.org/10.1111/geb.12099) [1111/geb.12099](https://doi.org/10.1111/geb.12099)
- <span id="page-13-15"></span>Bender MG, Machado GR, de Azevedo Silva PJ et al (2014) Local ecological knowledge and scientifc data reveal overexploitation by multigear artisanal fsheries in the Southwestern Atlantic. PLoS ONE 9(10):e110332. [https://](https://doi.org/10.1371/journal.pone.0110332) [doi.org/10.1371/journal.pone.0110332](https://doi.org/10.1371/journal.pone.0110332)
- <span id="page-13-8"></span>Bohnsack JA, Bannerot SP (1986) A stationary visual census technique for quantitatively assessing community structure of coral reef fshes. NOAA Technical Report NMFS  $41:1-15$
- <span id="page-13-4"></span>Brokovich E, Baranes A, Goren M (2006) Habitat structure determines coral reef fsh assemblages at the northern tip of the Red Sea. Ecol Indic 6:494–507. [https://doi.org/10.](https://doi.org/10.1016/j.ecolind.2005.07.002) [1016/j.ecolind.2005.07.002](https://doi.org/10.1016/j.ecolind.2005.07.002)
- <span id="page-13-1"></span>Carleton JH, Sammarco PW (1987) Efects of substratum irregularity on success of coral settlement: quantifcation by comparative geomorphological techniques. Bull Mar Sci 40:85–98
- <span id="page-13-7"></span>Castro B, Miranda L, Silva L et al (2008) Processos físicos: hidrografa, circulação e transporte. Oceanografa de um Ecossistema Subtropical: Plataforma de São Sebastião, SP, EDUSP, São Paulo
- <span id="page-14-29"></span>Cinner JE, Maire E, Huchery C, MacNeil MA, Graham NAJ, Mora C et al (2018) Gravity of human impacts mediates coral reef conservation gains. Proc Natl Acad Sci USA 115:E6116– E6125. <https://doi.org/10.1073/pnas.1708001115>
- <span id="page-14-25"></span>Cordeiro C, Mendes T, Harborne A, Ferreira C (2016) Spatial distribution of nominally herbivorous fshes across environmental gradients on Brazilian rocky reefs. J Fish Biol 89:939–958. <https://doi.org/10.1111/jfb.12849>
- <span id="page-14-6"></span>Cvitanovic C, Hoey AS (2010) Benthic community composition infuences within-habitat variation in macroalgal browsing on the Great Barrier Reef. Mar Freshwater Res 61:999–1005. <https://doi.org/10.1071/MF09168>
- <span id="page-14-2"></span>Darling ES, Graham NA, Januchowski-Hartley FA et al (2017) Relationships between structural complexity, coral traits, and reef fsh assemblages. Coral Reefs 36:561–575. <https://doi.org/10.1007/s00338-017-1539-z>
- <span id="page-14-21"></span>Daros FA, Bueno LS, Vilar CC et al (2012) Checklist of rocky reef fshes from the Currais Archipelago and Itacolomis Island, Paraná state, Brazil. Check List 8:349–354. [https://](https://doi.org/10.15560/8.3.349) [doi.org/10.15560/8.3.349](https://doi.org/10.15560/8.3.349)
- <span id="page-14-11"></span>de Freitas JEP, de Araújo ME, da Lotufo TM, C, (2019) Composition and structure of the ichthyofauna in a marine protected area in the western equatorial Atlantic: a baseline to support conservation management. Reg Stud Mar Sci 25:100488.<https://doi.org/10.1016/j.rsma.2018.100488>
- <span id="page-14-15"></span>De'Ath G (2002) Multivariate regression trees: a new technique for modeling species-environment relationships. Ecol 83:1105–1117.<https://doi.org/10.2307/3071917>
- <span id="page-14-17"></span>Elith J, Leathwick JR, Hastie T (2008) A working guide to boosted regression trees. J Anim Ecol 77:802–813. [https://](https://doi.org/10.1111/j.1365-2656.2008.01390.x) [doi.org/10.1111/j.1365-2656.2008.01390.x](https://doi.org/10.1111/j.1365-2656.2008.01390.x)
- <span id="page-14-4"></span>Endo CAK, Gherardi DFM, Pezzi LP, Lima LN (2019) Low connectivity compromises the conservation of reef fshes by marine protected areas in the tropical South Atlantic. Sci Rep 9:8634. [https://doi.org/10.1038/](https://doi.org/10.1038/s41598-019-45042-0) [s41598-019-45042-0](https://doi.org/10.1038/s41598-019-45042-0)
- <span id="page-14-8"></span>Ferrari R, Malcolm HA, Byrne M et al (2018) Habitat structural complexity metrics improve predictions of fsh abundance and distribution. Ecography 41:1077–1091. <https://doi.org/10.1111/ecog.02580>
- <span id="page-14-20"></span>Ferreira CEL, Gonçalves JEA, Coutinho R, Peret AC (1998) Herbivory by the dusky damselfsh *Stegastes fuscus* (Cuvier, 1830) in a tropical rocky shore: efects on the benthic community. J Exp Mar Bio Ecol 229(2):241– 264. [https://doi.org/10.1016/S0022-0981\(98\)00056-2](https://doi.org/10.1016/S0022-0981(98)00056-2)
- <span id="page-14-22"></span>Ferreira C, Floeter S, Gasparini J et al (2004) Trophic structure patterns of Brazilian reef fshes: a latitudinal comparison. J Biogeogr 31:1093–1106
- <span id="page-14-0"></span>Floeter SR, Guimaraes RZ, Rocha LA et al (2001) Geographic variation in reef-fsh assemblages along the Brazilian coast. Glob Ecol Biogeogr 10:423–431
- <span id="page-14-26"></span>Fonseca MS, Araújo FG, Teixeira-Neves TP et al (2021) Drivers of distribution of the parrotfsh *Sparisoma frondosum* (agassiz, 1831) in Southwest Atlantic rocky reefs: insights for management and conservation. Ocean Coast Manage 209:105642. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.ocecoaman.2021.105642) [ocecoaman.2021.105642](https://doi.org/10.1016/j.ocecoaman.2021.105642)
- <span id="page-14-19"></span>Francini-Filho RB, Moura RL (2008) Dynamics of fsh assemblages on coral reefs subjected to diferent management regimes in the Abrolhos Bank, eastern Brazil.

Aquat Conserv Mar Freshw Ecosyst 18:1166–1179. <https://doi.org/10.1002/aqc.966>

- <span id="page-14-5"></span>Friedlander AM, Parrish JD (1998) Habitat characteristics afecting fsh assemblages on a Hawaiian coral reef. J Exp Mar Bio Ecol 224:1–30. [https://doi.org/10.1016/](https://doi.org/10.1016/S0022-0981(97)00164-0) [S0022-0981\(97\)00164-0](https://doi.org/10.1016/S0022-0981(97)00164-0)
- <span id="page-14-1"></span>Friedlander A, Brown E, Jokiel P et al (2003) Efects of habitat, wave exposure, and marine protected area status on coral reef fsh assemblages in the Hawaiian archipelago. Coral Reefs 22:291–305. [https://doi.org/10.1007/](https://doi.org/10.1007/s00338-003-0317-2) [s00338-003-0317-2](https://doi.org/10.1007/s00338-003-0317-2)
- <span id="page-14-14"></span>Froese R. Pauly D (2020). FishBase. www. fshbase.org. Accessed 26 June 2020.
- <span id="page-14-7"></span>García-Charton JA, Pérez-Ruzafa Á, Sánchez-Jerez P et al (2004) Multi-scale spatial heterogeneity, habitat structure, and the efect of marine reserves on Western Mediterranean rocky reef fish assemblages.<br>Mar Biol 144:161-182. https://doi.org/10.1007/ 144:161–182. [https://doi.org/10.1007/](https://doi.org/10.1007/s00227-003-1170-0) [s00227-003-1170-0](https://doi.org/10.1007/s00227-003-1170-0)
- <span id="page-14-10"></span>Gibran FZ, Moura RL (2012) The structure of rocky reef fsh assemblages across a nearshore to coastal islands' gradient in Southeastern Brazil. Neotrop Ichthyol 10:369–382. <https://doi.org/10.1590/S1679-62252012005000013>
- <span id="page-14-27"></span>Giglio VJ, Moura RL, Gibran FZ et al (2019) Do managers and stakeholders have congruent perceptions on marine protected area management efectiveness? Ocean Coast Manag 179:104865. [https://doi.org/10.1016/j.ocecoaman.](https://doi.org/10.1016/j.ocecoaman.2019.104865) [2019.104865](https://doi.org/10.1016/j.ocecoaman.2019.104865)
- <span id="page-14-28"></span>Giglio VJ, Marconi M, Pereira-Filho GH, Leite KL, Figueroa AC, Motta FS (2022) Scuba divers' behavior and satisfaction in a new marine protected area: lessons from the implementation of a best practices program. Ocean Coast Manag 220:10691. [https://doi.org/10.1016/j.ocecoaman.](https://doi.org/10.1016/j.ocecoaman.2022.10609) [2022.10609](https://doi.org/10.1016/j.ocecoaman.2022.10609)
- <span id="page-14-16"></span>Greenwell B, Boehmke B, Cunningham J et al (2019) Package "gbm" R package version 2(5):37–40
- <span id="page-14-3"></span>Haupt PW, Lombard AT, Goodman PS, Harris JM (2017) Accounting for spatiotemporal dynamics in conservation planning for coastal fsh in KwaZulu-Natal, South Africa. Biol Conserv 209:289–303. [https://doi.org/10.](https://doi.org/10.1016/j.biocon.2017.02.009) [1016/j.biocon.2017.02.009](https://doi.org/10.1016/j.biocon.2017.02.009)
- <span id="page-14-12"></span>Hobson ES (1965) Diurnal-nocturnal activity of some inshore fshes in the Gulf of California. Copeia 1965(3):291–302.<https://doi.org/10.2307/1440790>
- <span id="page-14-9"></span>ICMBio (2017) Plano de manejo da Estação Ecológica Tupinambás e Refúgio de Vida Silvestre do Arquipélago de Alcatrazes. ICMBIo, Brasília
- <span id="page-14-18"></span>Imoto RD, Carneiro MH, Ávila-da-Silva AO (2016) Spatial patterns of fshing feets on the Southeastern Brazilian Bight. Lat Am J Aquat Res 44:1005–1018. [https://doi.](https://doi.org/10.3856/vol44-issue5-fulltext-1) [org/10.3856/vol44-issue5-fulltext-1](https://doi.org/10.3856/vol44-issue5-fulltext-1)
- <span id="page-14-13"></span>IUCN (2023) The IUCN Red List of Threatened Species. <https://www.iucnredlist.org/>. Accessed 12 Jan 2023.
- <span id="page-14-24"></span>Lewis A (1997) Recruitment and post-recruit immigration affect the local population size of coral reef fshes. Coral Reefs 16:139–149.<https://doi.org/10.1007/s003380050068>
- <span id="page-14-23"></span>Liuzzi MG, Gappa JL (2008) Macrofaunal assemblages associated with coralline turf: species turnover and changes in structure at diferent spatial scales. Mar Ecol Prog Ser 363:147–156. <https://doi.org/10.3354/meps07449>
- <span id="page-15-21"></span>Luiz OJ, Mendes TC, Barneche DR et al (2015) Community structure of reef fshes on a remote oceanic island (St Peter and St Paul's Archipelago, equatorial Atlantic): the relative infuence of abiotic and biotic variables. Mar Fresh Res 66:739–749. [https://doi.org/10.1071/](https://doi.org/10.1071/MF14150) [MF14150](https://doi.org/10.1071/MF14150)
- <span id="page-15-22"></span>Luiz Jr OJ, Carvalho-Filho A, Ferreira CE, et al (2008) The reef fsh assemblage of the Laje de Santos Marine State Park, Southwestern Atlantic: annotated checklist with comments on abundance, distribution, trophic structure, symbiotic associations, and conservation. Zootaxa 1807:1–25.<https://doi.org/10.11646/zootaxa.1807.1.1>
- <span id="page-15-1"></span>Maia HA, Morais RA, Quimbayo JP et al (2018) Spatial patterns and drivers of fsh and benthic reef communities at São Tomé Island. Tropical Eastern Atlantic Mar Ecol 39:e12520. <https://doi.org/10.1111/maec.12520>
- <span id="page-15-4"></span>Mallela J (2018) The infuence of micro-topography and external bioerosion on coral-reef-building organisms: recruitment, community composition and carbonate production over time. Coral Reefs 37:227–237. [https://doi.org/10.](https://doi.org/10.1007/s00338-017-1650-1) [1007/s00338-017-1650-1](https://doi.org/10.1007/s00338-017-1650-1)
- <span id="page-15-27"></span>Marconi M, Giglio VJ, Pereira Filho GH, Motta FS (2020) Does quality of scuba diving experience vary according to the context and management regime of marine protected areas? Ocean Coast Manage 194:105246. [https://doi.org/](https://doi.org/10.1016/j.ocecoaman.2020.105246) [10.1016/j.ocecoaman.2020.105246](https://doi.org/10.1016/j.ocecoaman.2020.105246)
- <span id="page-15-3"></span>Matheus Z, Francini-Filho RB, Pereira-Filho GH et al (2019) Benthic reef assemblages of the Fernando de Noronha Archipelago, tropical South-west Atlantic: effects of depth, wave exposure and cross-shelf positioning. PLoS ONE 14:e0210664.<https://doi.org/10.1371/journal.pone.0210664>
- <span id="page-15-12"></span>Minte-Vera CV, Moura RL, Francini-Filho RB (2008) Nested sampling: an improved visual-census technique for studying reef fsh assemblages. Mar Eco Prog Ser 367:283– 293. <https://doi.org/10.3354/meps07511>
- <span id="page-15-19"></span>Mora C, Aburto-Oropeza O, Bocos AA et al (2011) Global human footprint on the linkage between biodiversity and ecosystem functioning in reef fshes. PLoS Biol 9:e1000606.<https://doi.org/10.1371/journal.pbio.1000606>
- <span id="page-15-26"></span>Morais R, Ferreira C, Floeter S (2017) Spatial patterns of fish standing biomass across Brazilian reefs. J Fish Biol 91(6):1642–1667.<https://doi.org/10.1111/jfb.13482>
- <span id="page-15-13"></span>Motta FS, Moura RL, Neves LM et al (2021) Efects of marine protected areas under diferent management regimes in a hot spot of biodiversity and cumulative impacts from SW Atlantic. Reg Stud Mar Sci 47:101951. [https://doi.org/10.](https://doi.org/10.1016/j.rsma.2021.101951) [1016/j.rsma.2021.101951](https://doi.org/10.1016/j.rsma.2021.101951)
- <span id="page-15-15"></span>Moura RL, Francini-Filho RB (2005) Reef and shore fshes of the Abrolhos region, Brazil. In: Dutra GF, Allen GR, Werner T, McKenna SA (eds) A rapid marine biodiversity assessment of the Abrolhos Bank, Bahia, Brazil, 1st edn. Conservation International, Washington DC, pp 40–55
- <span id="page-15-10"></span>Nanami A (2022) Co-occurrence of seagrass vegetation and coral colonies supports unique fsh assemblages: a microhabitat-scale perspective. PeerJ 10:e14466. [https://doi.org/](https://doi.org/10.7717/peerj.14466) [10.7717/peerj.14466](https://doi.org/10.7717/peerj.14466)
- <span id="page-15-9"></span>Neves LM, Teixeira-Neves TP, Pereira-Filho GH, Araújo FG (2016) The farther the better: effects of multiple environmental variables on reef fsh assemblages along a distance gradient from river infuences. PLoS ONE 11:e0166679. <https://doi.org/10.1371/journal.pone.0166679>
- <span id="page-15-6"></span>O'Connor MI, Bruno JF, Gaines SD et al (2007) Temperature control of larval dispersal and the implications for marine ecology, evolution, and conservation. Proc Natl Acad Sci USA 104:1266–1271.<https://doi.org/10.1073/pnas.0603422104>
- <span id="page-15-0"></span>Parravicini V, Kulbicki M, Bellwood D et al (2013) Global patterns and predictors of tropical reef fsh species richness. Ecography 36:1254–1262. [https://doi.org/10.1111/j.1600-](https://doi.org/10.1111/j.1600-0587.2013.00291.x) [0587.2013.00291.x](https://doi.org/10.1111/j.1600-0587.2013.00291.x)
- <span id="page-15-24"></span>Pereira PHC, Ferreira BP (2013) Effects of life phase and schooling patterns on the foraging behaviour of coral-reef fshes from the genus *Haemulon*. J Fish Biol 82:1226– 1238.<https://doi.org/10.1111/jfb.12054>
- <span id="page-15-25"></span>Pereira PHC, Barros B, Zemoi R, Ferreira BP (2015) Ontogenetic diet changes and food partitioning of *Haemulon* spp. coral reef fshes, with a review of the genus diet. Rev Fish Biol Fish 25:245–260. [https://doi.org/10.1007/](https://doi.org/10.1007/s11160-014-9378-2) [s11160-014-9378-2](https://doi.org/10.1007/s11160-014-9378-2)
- <span id="page-15-2"></span>Pereira PHC, Macedo CH, Nunes JACC et al (2018) Effects of depth on reef fsh communities: Insights of a "deep refuge hypothesis" from Southwestern Atlantic reefs. PLoS ONE 13:e0203072. <https://doi.org/10.1371/journal.pone.0203072>
- <span id="page-15-20"></span>Pereira-Filho GH, Amado-Filho GM, Guimarães S et al (2011) Reef fsh and benthic assemblages of the Trindade and Martin Vaz Island group. Southwestern Atlantic Braz J Oceanogr 59(3):201–212
- <span id="page-15-16"></span>Pinheiro HT, Rocha LA, Macieira RM et al (2018) Southwestern Atlantic reef fshes: zoogeographical patterns and ecological drivers reveal a secondary biodiversity centre in the Atlantic Ocean. Divers Distrib 24:951–965. [https://](https://doi.org/10.1111/ddi.12729) [doi.org/10.1111/ddi.12729](https://doi.org/10.1111/ddi.12729)
- <span id="page-15-11"></span>Pires-Vanin A, Rossi-Wongtschowski C, Aidar E et al (1993) Estrutura e função do ecossistema de plataforma continental do Atlântico Sul brasileiro: síntese dos resultados. Publicação Especial Do Instituto Oceanográfco 10:217–231
- <span id="page-15-8"></span>Quimbayo JP, Dias MS, Kulbicki M et al (2019) Determinants of reef fsh assemblages in tropical oceanic islands. Ecography 42:77–87. <https://doi.org/10.1111/ecog.03506>
- <span id="page-15-23"></span>Quimbayo JP, Silva FC, Mendes TC et al (2021) Life-history traits, geographical range, and conservation aspects of reef fshes from the Atlantic and Eastern Pacifc. Ecology 102:e03298.<https://doi.org/10.1002/ecy.3298>
- <span id="page-15-17"></span>R Core Team (2011) R: a language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>. Accessed 20 Apr 2022
- <span id="page-15-7"></span>Richards BL, Williams ID, Vetter OJ, Williams GJ (2012) Environmental factors afecting large-bodied coral reef fsh assemblages in the Mariana Archipelago. PLoS ONE 7:e31374. <https://doi.org/10.1371/journal.pone.0031374>
- <span id="page-15-18"></span>Rolim FA, Langlois T, Rodrigues PF et al (2019) Network of small no-take marine reserves reveals greater abundance and body size of fsheries target species. PLoS ONE 14:e0204970. <https://doi.org/10.1371/journal.pone.0204970>
- <span id="page-15-14"></span>Rolim FA, Gibran FZ, Neves LM et al (2024) Is a recent large expansion of a no-take reserve sufficient to change the reef fsh assemblage? A before-after assessment in Alcatrazes Wildlife Refuge. Southwestern Atlantic Ocean Coast Manag 250:107043. [https://doi.org/10.1016/j.oceco](https://doi.org/10.1016/j.ocecoaman.2024.107043) [aman.2024.107043](https://doi.org/10.1016/j.ocecoaman.2024.107043)
- <span id="page-15-5"></span>Russ GR, Rizzari JR, Abesamis RA, Alcala AC (2021) Coral cover a stronger driver of reef fsh trophic biomass than

fshing. Ecol Appl 31:e02224. [https://doi.org/10.1002/eap.](https://doi.org/10.1002/eap.2224) [2224](https://doi.org/10.1002/eap.2224)

<span id="page-16-2"></span>Sale PF (2013) The ecology of fshes on coral reefs. Elsevier

- <span id="page-16-10"></span>Sánchez-Caballero CA, Borges-Souza JM, Ferse SCA (2019) Rocky reef fsh community composition remains stable throughout seasons and El Niño/La Niña events in the southern Gulf of California. J Sea Res 146:55-62. [https://](https://doi.org/10.1016/j.seares.2019.01.008) [doi.org/10.1016/j.seares.2019.01.008](https://doi.org/10.1016/j.seares.2019.01.008)
- <span id="page-16-7"></span>Silva MB, Rosa RS, Menezes R, Francini-Filho RB (2021) Changes in reef fsh assemblages in a cross-shelf euphotic-mesophotic gradient in tropical SW Atlantic. Estuar Coast Shelf Sci 259:107465. [https://doi.org/10.](https://doi.org/10.1016/j.ecss.2021.107465) [1016/j.ecss.2021.107465](https://doi.org/10.1016/j.ecss.2021.107465)
- <span id="page-16-9"></span>Silvano RAM, Güth AZ (2006) Diet and feeding behavior of *Kyphosus* spp (Kyphosidae) in a Brazilian subtropical reef. Braz Arch Biol Technology 49(4):623–629. [https://](https://doi.org/10.1590/S1516-89132006000500012) [doi.org/10.1590/S1516-89132006000500012](https://doi.org/10.1590/S1516-89132006000500012)
- <span id="page-16-6"></span>Souza GRS de, Gadig OBF, Motta FS, et al (2018) Reef fshes of the Anchieta Island State Park, Southwestern Atlantic, Brazil. Biota Neotrop 18(1). [https://doi.org/10.1590/](https://doi.org/10.1590/1676-0611-BN-2017-0380) [1676-0611-BN-2017-0380](https://doi.org/10.1590/1676-0611-BN-2017-0380)
- <span id="page-16-4"></span>Takase LS, Stein LP, Hoff NT, Siegle E (2021) Wave climate and power distribution around a rocky island: Alcatrazes, Brazil. Ocean Coast Res 69. [https://doi.org/10.1590/2675-](https://doi.org/10.1590/2675-2824069.20-009lst) [2824069.20-009lst](https://doi.org/10.1590/2675-2824069.20-009lst)
- <span id="page-16-0"></span>Teixeira-Neves TP, Neves LM, Araújo FG (2015) Hierarchizing biological, physical and anthropogenic factors influencing the structure of fish assemblages along tropical rocky shores in Brazil. Environ

Biol Fish 98:1645–1657. [https://doi.org/10.1007/](https://doi.org/10.1007/s10641-015-0390-8) [s10641-015-0390-8](https://doi.org/10.1007/s10641-015-0390-8)

- <span id="page-16-5"></span>Trygonis V, Sini M (2012) photoQuad: a dedicated seabed image processing software, and a comparative error analysis of four photoquadrat methods. J Exp Mar Biol Ecol 424:99–108. <https://doi.org/10.1016/j.jembe.2012.04.018>
- <span id="page-16-8"></span>Von Glasow R, Jickells TD, Baklanov A et al (2013) Megacities and large urban agglomerations in the coastal zone: interactions between atmosphere, land, and marine ecosystems. Ambio 42:13–28. <https://doi.org/10.1007/s13280-012-0343-9>
- <span id="page-16-1"></span>Williams ID, Baum JK, Heenan A et al (2015) Human, oceanographic and habitat drivers of central and western Pacifc coral reef fsh assemblages. PLoS ONE 10:e0120516. <https://doi.org/10.1371/journal.pone.0120516>
- <span id="page-16-3"></span>Wilson S, Graham NAJ, Polunin NVC (2007) Appraisal of visual assessments of habitat complexity and benthic composition on coral reefs. Mar Biol 151:1069–1076. [https://doi.](https://doi.org/10.1007/s00227-006-0538-3) [org/10.1007/s00227-006-0538-3](https://doi.org/10.1007/s00227-006-0538-3)

**Publisher's Note** Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Springer Nature or its licensor (e.g. a society or other partner) holds exclusive rights to this article under a publishing agreement with the author(s) or other rightsholder(s); author self-archiving of the accepted manuscript version of this article is solely governed by the terms of such publishing agreement and applicable law.